

ACUTE TOXICITY OF ZINC AND COPPER FOR RAINBOW TROUT (*Onchorhynchus mykiss*)

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Abstract: Anchovy oil is a very suitable supplementary ingredient for fish feeds due to the essential fatty acid composition. The acute toxicity of zinc and copper ions for rainbow trout (*Oncorhynchus mykiss* Walbaum 1792) were evaluated by static bioassays. The average weight and length of fish used in the zinc experiments were $3,02 \pm 0.21$ g and 6.52 ± 0.12 cm, respectively, while the tests with copper ions were performed with larger fish (7.12 ± 0.60 g and 7.89 ± 0.12 cm). Temperature, dissolved O₂, pH and ammonia were measured daily, and the average values were $14.62 \pm 0.41^\circ\text{C}$, 7.49 ± 0.15 mg/l O₂, 7.48 ± 0.12 and 0.013 ± 0.002 mg/l NH₃-N, respectively (total hardness of 249.6 mg/l CaCO₃). Chemically pure salts of zinc chloride (ZnCl₂) and copper sulphate (CuSO₄ 5H₂O) dissolved in distilled water were used as toxicants. Eight zinc ion concentrations with a control group and 8 copper ion concentrations with a control group were prepared. The LT₅₀ (lethality time for 50%) and 96-h LC₅₀ (lethal concentration for 50%) values were calculated. The LC₅₀ values of zinc and copper ions for rainbow trout were found to be 12.88 and 0.094 mg/l, respectively. Survival time decreased with increasing concentrations of zinc and copper ions. Copper ion concentrations were found to be more toxic than zinc ion concentrations for rainbow trout.

Keywords: Zinc, copper, LT₅₀, LC₅₀, mortality, rainbow trout

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Introduction

The harmful effects of heavy metals on rainbow trout have been studied by various authors (Taylor et al., 2000; Hansen et al., 2002b; Bagdonas and Vosyliene 2006). Include alterations in metabolism and specific physiological functions (Glover and Hogstrand, 2003) leading to irreversible damages in behavior (Petrauskienė, 1999), growth, and reproduction (Kleerekoper, 1976; Nieboer and Richardson, 1980; Mance, 1987; Depledge et al., 1994; Phillips and Rainbow, 1994).

The effects of sublethal doses of zinc and copper on fish behavior were also studied and complicated responses found in their biological systems (Suresh et al., 1993).

Toxic substances dissolved in water increase often the sensitivity of aquatic organisms to temperature variations, changes in dissolved O₂ and vice-versa. Also the growth performance can be impaired and reproduction capacity can be reduced. Metabolic effects of heavy metal exposure (i.e. oxygen consumption) can be explained by the accumulation of heavy metals on the gill surface impairing O₂ diffusion capacity (Mutluay and Demirak, 1996; Svecevičius, 1999; Taylor et al., 2000 -2002; Bagdonas and Vosyliene 2006).

Heavy metals can be taken up by aquatic organisms via several routes (a) directly via the body surface or surface of the respiratory organs, (b) via feed, or (c) by a combination of them (Phillips and Rainbow, 1994). However which route is more important depends on environmental circumstances and has not always been properly documented (Depledge et al., 1994). Aquatic toxicology is also concerned with effects of the concentrations of the chemical substances occurring in natural waters, sediments and food. The aquatic environment is complex and exhibits a changeable structure. These changing conditions affect the chemical reactions of substances and pollutants (Cairns and Mount, 1990; Forbes and Forbes, 1994).

Most of the above mentioned heavy metals studies have determined the LT₅₀ and LC₅₀ values for various fish species but also assessed the accumulation in fish tissues illustrate in also that effects change with changing environmental conditions. Keeping these facts in our minds, the aim of the study was to find out the effects of zinc and copper ions for the rainbow trout in laboratory environmental conditions.

Material and Methods

Rainbow trout (*Oncorhynchus mykiss*) was purchased from a private fish farm and transported to the laboratory and placed in tanks (500 L), which the stocking density was 2000 fish/m³. The fish were fed on a commercial diet, containing 52% crude protein, 14% crude lipid and 19 KJ gross energy/g feed (Sibal A.Ş., Sinop, Turkey). The acclimation to test conditions lasted for 10 days. The same diet was used during the whole experiment. Tap water was aerated for 48 hours to remove chlorine. Half of the water in the stocking tanks was changed every other day with well-aerated water containing the same concentration of heavy metal.

Chemically pure salts of zinc chloride (ZnCl₂) and copper sulphate (CuSO₄ 5H₂O) dissolved in distilled water were used as toxicants. The final concentrations were recalculated according to the amount of zinc and copper ions.

The average weight and length of fish used in the zinc experiments range from 2.81 ± 0.24 to 3.21 ± 0.26 g and 6.37 ± 0.12 to 6.63 ± 0.11 cm respectively, while the weight and length for copper experiments range from 6.71 ± 0.78 to 7.46 ± 0.63 g and 7.06 ± 0.32 to 8.41 ± 0.24 cm respectively. Temperature, dissolved O₂, pH, total hardness and ammonia were measured daily, and the average values were 14.62 ± 0.41 °C, 7.49 ± 0.15 mg/l O₂, 7.48 ± 0.12, 249.56 mg/l as CaCO₃ (Ca⁺⁺ + Mg⁺⁺ = 85.86 + 8.83 mg CaCO₃ / l) and 0.013 ± 0.002 mg/l NH₃-N, respectively.

Two separate experiments were conducted to determine the lethal dosages of zinc and copper ions. The experiments were designed as 3 replicates in the tanks containing 100 L of water, each replicate had 8 exposure groups as well as one control and each test group was performed with ten fish.

Experiment I contained 8 zinc ion concentrations (5, 8, 10, 13, 15, 19, 23 and 27 mg/l) besides the control (not contain zinc) while Experiment II contained 8 copper ion concentrations (0.01, 0.025, 0.05, 0.075, 0.1, 0.5, 1 and 2 mg/l) besides the control (not contain copper). Concentration were determined by analysing the samples through atomic absorption spectrometer.

Observations were made after 15, 30 min, 1, 2, 4, 8 and 12 h on the first day, while follow-up observations were conducted after 24, 48, 72, 96 had 120 h. Death or abnormality in swimming

behaviour of fish were noted. The tanks were checked daily; pH and dissolved O₂ values were measured throughout the experimental period. Fish were not fed for 1 day before the start of experiments to the end of the 120-h experimental period. Thus, the volume of waste matter was minimized for not affect the fish conditions. Death was diagnosed either by lack of movement of the operculum or inactivity in swimming behaviour (Ünsal, 1998).

The terms LC₅₀ and LT₅₀ are in accordance with Lloyd (1992) and LC₅₀ concentration values were analyzed by Probite Analysis (Finney, 1971). Data on mortalities recorded in the three replicates for each concentration were pooled. The mortality percentage of each concentration was verified using Abbott's formula (Anonymous, 1976). Weighted regression analysis based on probate (transformed percentage mortality) against log-dose was calculated for each metal independently, and considering these calculations for the lethal concentrations (LC₅₀) and fiducially limits (FL) was determined.

Results and Discussion

The mortality rates increased with increasing zinc and copper ions concentrations as shown in Figures 1 and 2. No mortality occurred in the control groups. Also no fish died within the first 4 h in the concentrations up to 13 mg/l Zn whereas mortality was occurred in the concentrations from

15 to 27 mg/l. In the lowest zinc concentration, first mortality was seen after 48 hours. No mortality occurred during the first two hour of exposure in any replicate of any of the copper ion concentrations employed. Figure 2 shows that mortality after 4 h exposure reached 3.3% in the concentration of 2 mg/l Cu. In addition, the death rates were 3.3 and 6.7% in the 0.5-1 mg/l Cu solutions respectively after 8 h exposure, and a death rate of 6.7% in the concentration of 0.1 mg/l Cu was recorded at the end of 12 h exposure. None of the fish in the control group died during the experimental period. In the highest concentration (2 mg/l Cu), the first fish died after 4 h exposure, and all of the fish died at the end of 96 h exposure.

Mortalities in the highest three zinc levels were categorized in the same statistical group whereas mortalities in the lowest two zinc levels and the control group were categorized as another group and the rest formed the last group. When these three groups were compared, the mortality percentages were found to be significantly different from each other as shown in Table 1 ($p < 0.05$). Mortalities in the lowest four copper levels and the control group were categorized in the same statistical group and the rest formed another group. When these groups were compared, the results were significantly different (Table 1, $p < 0.05$).

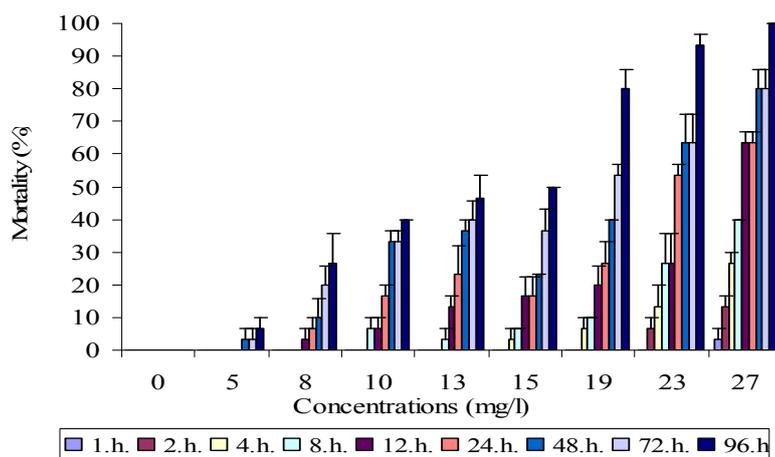


Figure 1. Percentage mortality at different zinc ion concentrations for different exposure times (h) Data present means \pm standard error based on 3 replicates with 10 fish each.

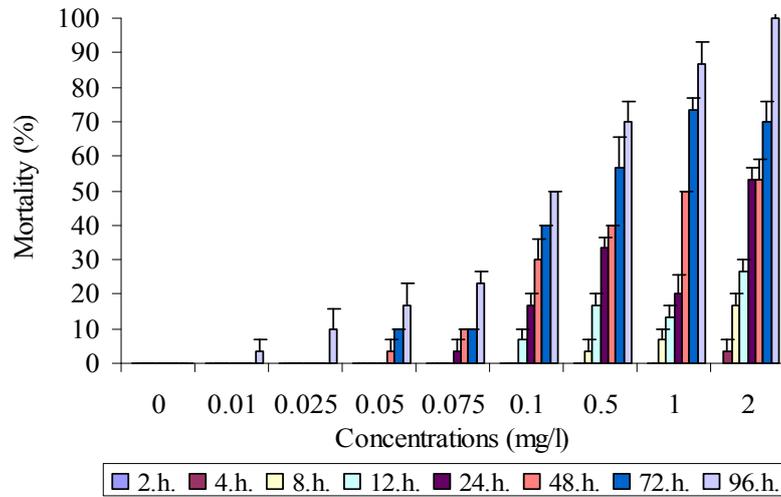


Figure 2. Percentage mortality at different copper ion concentrations for different exposure times (h). Data present means \pm standard error based on 3 replicates with 10 fish each.

Table 1. Mortality (%) in different zinc and copper ion concentrations at the end of 96h exposure experiment. Data present means \pm standard error based on 3 replicates with 10 fish each

Concentration (Zn mg/l)	% Mortality \pm SEM Zn	Concentration (Cu mg/l)	% Mortality \pm SEM Cu
Control	0 \pm 0.0 ^a	0	0 \pm 0.0 ^a
5	6.7 \pm 3.3 ^a	0.01	3.3 \pm 3.3 ^a
8	26.7 \pm 3.3 ^a	0.025	10 \pm 5.7 ^a
10	40 \pm 0.0 ^b	0.05	16.7 \pm 6.7 ^a
3	46.7 \pm 6.7 ^b	0.075	23.3 \pm 3.3 ^a
15	50 \pm 0.0 ^b	0.1	50 \pm 0.0 ^b
19	80 \pm 0 \pm 0.0 ^c	0.5	70 \pm 5.8 ^{bc}
23	93.3 \pm 3.3 ^c	1	86.7 \pm 6.7 ^c
27	100 \pm 0.0 ^c	2	100 \pm 6.7 ^c

The LT_{50} for rainbow trout in different zinc and copper ion concentrations are shown in Figures 3 and 4 respectively. Since 50% mortality was not recorded in the 96 h exposure, the time-span of the experiments was extended up to 120 hours. The LT_{50} values in 15 and 27 mg/l Zn concentrations were 96 and 10 h, respectively. The LT_{50} values in 0.1 and 2 mg/l Cu concentrations were 96 and 22 h, respectively. There is negative correlation between the LT_{50} values and the zinc and copper ion concentrations; when the zinc and copper ion concentrations levels decreased, LT_{50} values increased.

The mortality (%) in different zinc and copper ion concentrations within 96 h was evaluated as well. The results illustrated that 50 and 100% of fish died in the concentrations of (LC_{50}) 15 and 27 mg/l Zn solutions, respectively (Figure 5). For

copper ion concentrations, 50 and 100 % mortality was found in (LC_{50}) 0.1 and 2 mg/l Cu solutions, respectively (Figure 6).

The relationships both between Zn concentrations and mortality and between Cu concentrations and mortality were analyzed. The results in the basic correlation analysis illustrated a positive linear relationship as follows; $y=12.101x-22.2$ and $r=0.9835$ for Zn and $y=13.389x-26.944$ and $r=0.9694$ for Cu, respectively (Figure 5 and 6).

Lastly, the concentration values causing 50% mortality at the end of the 96-h period were analysed and the results were displayed in Table 2. LC_{50} was found only in the concentrations of 15 mg/l Zn and 0.1 mg/l Cu after 96 h exposure. These concentration values were analysed by Probit Analysis and LC_{50} value was calculated as 12.8 mg/l Zn and 0.094 mg/l Cu.

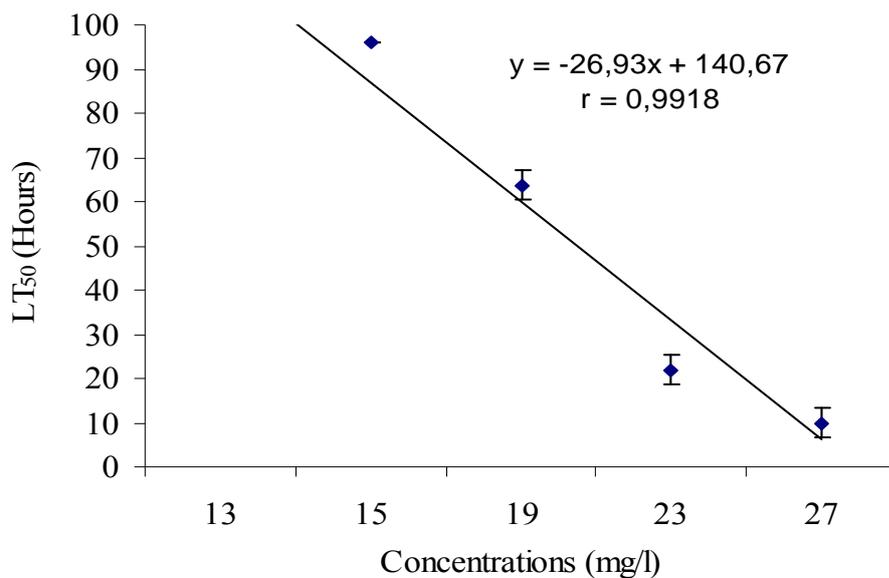


Figure 3. Time span for LT_{50} at different zinc ion concentrations at a given water hardness of 249.56 mg/l as $CaCO_3$. Data present means \pm standard error based on 3 replicates with 10 fish each.

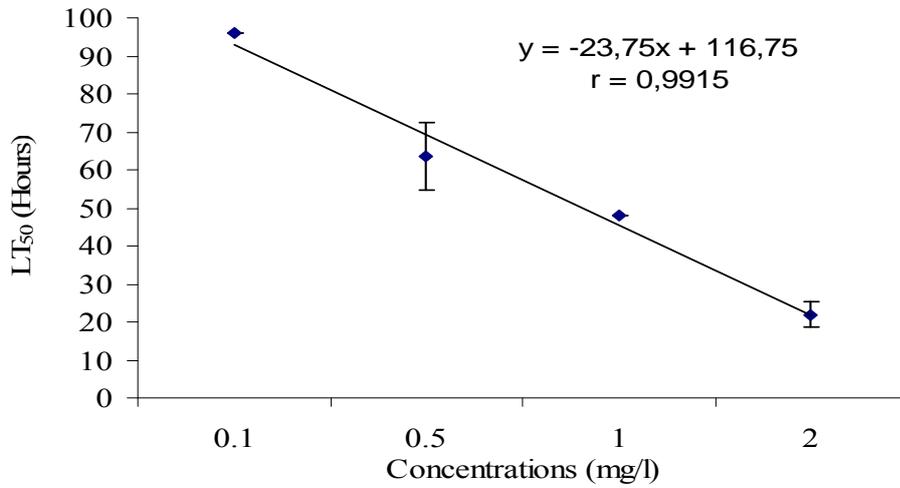


Figure 4. Time span for LT₅₀ at different copper ion concentrations at a given water hardness of 249.56 mg/l as CaCO₃. Data present means ± standard error based on 3 replicates with 10 fish each.

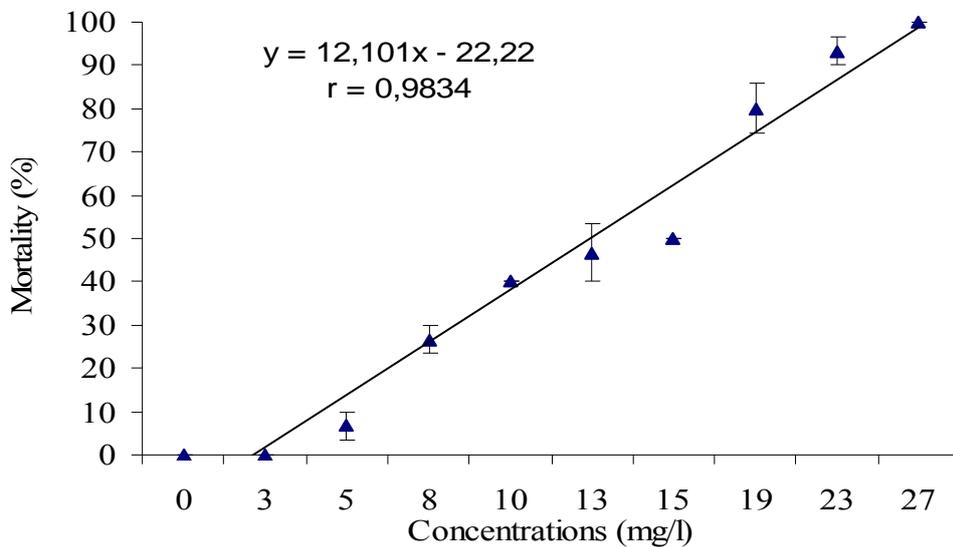


Figure 5. Mortality (%) in different zinc ion concentrations at the end of 96 h exposure experiment. Data present means ± standard error based on 3 replicates with 10 fish each.

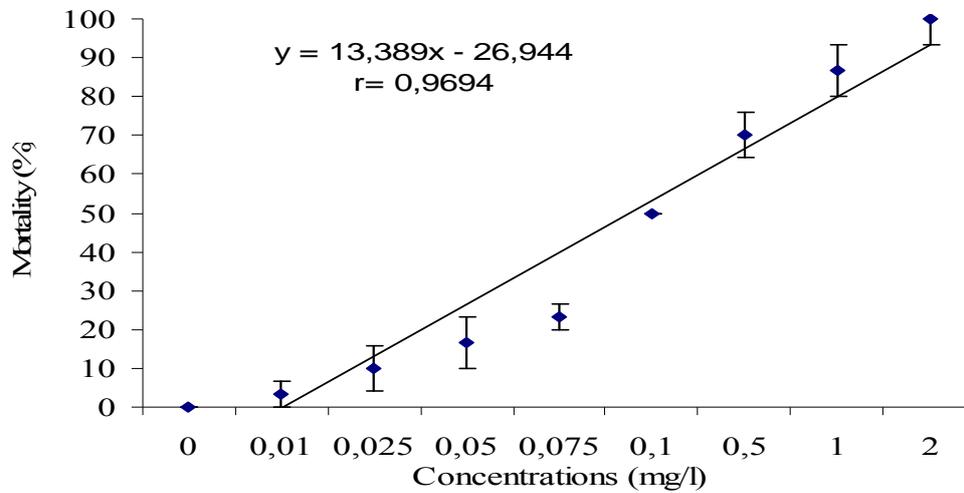


Figure 6. Mortality (%) in different copper ion concentrations at the end of 96 h exposure experiment. Data present means \pm standard error based on 3 replicates with 10 fish each.

Table 2. LC₅₀ value (mg/l) and slope (b) with 95% fiducially (FL) and 95% confidence (CL) limits, intercept (a) and χ^2 value of 96 h probate line for rainbow trout exposed to zinc and copper ions.

96 h	Zn (mg/l)	Cu (mg/l)
LC (95% FL)	12.8 (9.81 – 15.94)	0.094 (0.05 – 0.13)
b (95% CL)	2.10	3.02
a	3.01	7.93
χ^2	3.42	14.02

Several researchers reported different lethal dosages for zinc under different water conditions (Sprague and Ramsay, 1965; Goettl et al., 1976; Hale, 1977; Svecevičius, 1999; Bagdonas and Vosyliënė 2006). The comparison of the results of the present study with the above mentioned studies was illustrated in Table 3. The differences in LC₅₀ values might be caused by the different metal compounds used in the studies and environmental conditions in which the studies were applied. The toxicity limits of zinc in rainbow trout for water hardness of 10–500 mg/l as CaCO₃ and a pH value of 6 were ≤ 0.03 and ≤ 0.50 mg/l, respectively, and the effects in hard water were found to be lower than that of soft water (Lloyd, 1992), and differences may also be due to the

different ratios of carbonate hardness to total hardness. Roy and Campbell (1995) reported varying LC₅₀ values for young Atlantic salmon exposed to different Zn solution concentrations at different pH levels and the limits were reported to be 0.1–0.2 μ M. Hansen et al. (2002b) also reported various effects of zinc in rainbow trout at different hardness, pH and temperature levels, and noted a low level toxicity at high hardness and low pH values.

The results of this study indicated that mortality rate and time were influenced by the concentration levels of the heavy metals as well as the kind of metals used. Besides, it was found that there was a positive relationship between the mortality and concentration levels; when the con-

centration level increased, the mortality rate increased as well. However, there was a negative relationship between the mortality time and concentration level; when the concentration level increased, the mortality time decreased.

In the previous studies, different LC_{50} and LT_{50} values were detected (Table 3). The experiments of zinc solutions were conducted in conditions of total hardness of 249.6 mg/l as $CaCO_3$, temperature level of 12 °C and pH value of 7.5 Wong et al. (1977) and Lam et al. (1998) also conducted similar experiments using different fish species, namely common carp in the first and tilapia in the second study, and different heavy metal concentrations. There were some differences between our findings and the results of the above-mentioned researchers in the toxicity levels of zinc, possibly because of the different fish species used. This result clearly reflected that the toxicity levels of heavy metals varied depending on experimental fish species.

The time-spans for 50% mortality in different copper concentrations were analyzed. Schaeperclaus (1979) and Reichenbach-Klinke (1980) conducted experiments as well and found the lethal concentrations of copper to be 0.1 and 0.143 mg/l Cu, respectively. In another study, similar results were also observed (see Table 3; Chapman and McCrady, 1977). All of these results are very close to the results of our study (0.094 mg/l).

Several researchers reported different LC_{50} values for rainbow trout fed at different water conditions containing copper: an LC_{50} value of 0.01 mg/l as $CuSO_4$ at a hardness of 13 mg/l as $CaCO_3$ and pH value of 7.2 (Chapman and McCrady, 1977), high mortality rate and lower growth performance in water containing 144 µg/l Cu (Dixon and Hilton, 1985) and an LC_{50} value of 14 µg/l Cu (Marr et al., 1998). However, in the present study, the LC_{50} value of 0.094 mg/l Cu is lower than those mentioned above. This may be caused by the fish size used in the study, as reported by Howarth and Sprague (1978). Some contradictions in the LC_{50} values were also found between our results and those of Hansen et al. (2002a), supporting the varying effects of copper, depending on different water conditions, different fish sizes used and different copper salts in the experiments.

Brown and Dalton (1970) found a different LT_{50} level (2 d=48 h) from their experimental study using $CuSO_4$ in similar water hardness and

pH values (Table 3). This difference can be explained by the difference in LC_{50} values and fish species used. On the other hand, the 50% mortality in 96 h (LT_{50}) exposure for Cu toxicity obtained by Taylor et al. (2000) and Bagdonas and Vosyliene (2006) supports our experimental results (Table 3). A LC_{50} value of 0.75 mg/l was analyzed for $CuSO_4$ solution for 48 h at 17°C in *Salmo gairdneri* weighing 8.5 gr (Brown and Dalton, 1970), on the other hand, a LC_{50} value of 1 mg/l was determined for the same solution for 48 h at 12 °C in the same fish species weighing 6.71-7.46 g. This difference can be explained by different experimental water temperature at both studies.

Table 3 shows that LT_{50} values for Zn toxicity were analyzed as 4 days (96 h) in the present study, and also by Goettl et al. (1976). However, the LC_{50} values were different in our study and in Goettl's, probably as a result of the differences in water hardness, Zn salt solutions and fish species used. A LC_{50} value of 4.52 mg/l was determined for $ZnSO_4$ solution at the hardness of 312 mg/l as $CaCO_3$ in *Salmo gairdneri* (Goettl et al., 1976), on the other hand, in our study, a LC_{50} value of 12.9 mg/l was determined for the same solution at the hardness of 249.6 mg/l as $CaCO_3$ in the similar fish species. The increase in the LC_{50} value in our study can be explained by the decrease in water hardness (Table 3).

Conclusion

The results of this study clearly illustrated that the toxic effects of zinc and copper to fish, i.e. the LT_{50} and LC_{50} values varied according to water conditions, such as temperature, pH, hardness, dissolved O_2 , the size and species of fish as well as the type of zinc or copper salt.

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I assure you that whole activities involving experimental fish in the present study were conducted in accordance with national and institutional guidelines for the protection of human subjects and animal welfare.

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Table 3. Toxicity levels of zinc and copper ions concentration in freshwater fish species (h: hour, d: day)

<i>Fish</i>	Hardness (as pH mg/l CaCO ₃)	Temperature (°C)	Metal component	Concentration (mg/l LC ₅₀)	Time LT ₅₀	Reference
<i>Oncorhynchus tshawytscha</i> (1.4 g)	13 46 182 359	7.2 7.6 8.1 8.5	-	CuSO ₄ 0.01 mg/l 0.025 mg/l 0.09 mg/l 0.125 mg/l	4 d	Chapman and McCrady, 1977. (cited by Mance, 1987)
<i>Salmo gairdneri</i> (4 g) (8,5 g) (60 days)	125 240 82-132	7.8 7.4 6.4-8.3	10 17 -	CuSO ₄ CuSO ₄ ZnSO ₄ Cu(NO ₃) ₂ 0.2 mg/l 0.75 mg/l 4 mg/l 0.253 mg/l	4 d 2 d 4 d	Spear and Anderson, 1975. (cited by Mance, 1987) Brown and Dalton, 1970 Hale, 1977 (cited by Mance, 1987)
<i>Salmo gairdneri</i> (60 days) (7 cm)	82-132 22 312 312	6.4-8.3 - - -	- - - -	Zn(CH ₃ COO) ₂ ZnSO ₄ 1.19 mg/l 4.52 mg/l	4 4	Hale, 1977. Goettl et. al., 1976. (cited by Mance, 1987)
<i>Oncorhynchus mykiss</i>	30-360 (Alkalinity: 520 µg/l)	5-9	-	CuOH ⁺ and Cu ₂ OH ⁺² 20 µg/l	96 h	Howarth and Sprague, 1978.
<i>Oncorhynchus Mykiss</i>	270-300	7.9-8.1	10	ZnSO ₄ 3.79 mg/l	96 h	Svecevičius, 1999.
<i>Oncorhynchus mykiss</i>	120	8	-	Cu 100 µg/l	96 h	Taylor et. al., 2000.
<i>Oncorhynchus mykiss</i>	284	8	12	CuSO ₄ ZnSO ₄ 0.16 mg/l 0.95 mg/l	96 h	Bagdonas and Vosyliene, 2006
<i>Oncorhynchus mykiss</i>	249.56	7.46	12	CuSO ₄ 0.094 mg/l	96 h	This study
	249.56	7.46	12	ZnCl ₂ 12.88 mg/l	96 h	